







Association between exposure to air pollution and blood lipids in the general population of Spain

Sergio Valdés^{1,2}  | Viyey Doulatram-Gamgaram¹  | Cristina Maldonado-Araque^{1,2}  |
 Eva García-Escobar^{1,2}  | Sara García-Serrano^{1,2} | Wasima Oualla-Bachiri^{1,2} |
 Marta García-Vivanco³ | Juan Luis Garrido³ | Victoria Gil³ |
 Fernando Martín-Llorente³ | Alfonso Calle-Pascual^{2,4} | Luis Castaño^{2,5,6} |
 Elías Delgado⁷ | Edelmiro Menéndez⁷ | Josep Franch-Nadal^{2,8} |
 Sonia Gaztambide^{2,6,9} | Joan Girbés¹⁰  | F. Javier Chaves^{2,11} | José L. Galán-García¹² |
 Gabriel Aguilera-Venegas¹² | Joan Carles Vallvé^{2,13} | Núria Amigó^{2,14,15} |
 Montse Guardiola^{2,13} | Josep Ribalta^{2,13} | Gemma Rojo-Martínez^{1,2} 

¹Department of Endocrinology and Nutrition, Hospital Regional Universitario de Málaga/Universidad de Málaga, Instituto de Investigación Biomedica de Málaga-IBIMA, Málaga, Spain

²Centro de Investigación Biomédica en Red de Diabetes y Enfermedades Metabólicas Asociadas (CIBERDEM), Instituto de Salud Carlos III, Madrid, Spain

³Centro de Investigaciones Energéticas, Medioambientales y Tecnológicas (CIEMAT) – División de Contaminación Atmosférica, Madrid, Spain

⁴Department of Endocrinology and Nutrition and Instituto de Investigación Sanitaria University Hospital S. Carlos (IdISSC), Department Medicine II, Universidad Complutense (UCM), Madrid, Spain

⁵Hospital Universitario Cruces, BioCruces, UPV/EHU, Barakaldo, Spain

⁶Centro de Investigación Biomédica en Red de Enfermedades Raras (CIBERER), Instituto de Salud Carlos III, Madrid, Spain

⁷Department of Endocrinology and Nutrition, Hospital Universitario Central de Asturias/University of Oviedo, Instituto de Investigación Sanitaria del Principado de Asturias (ISPA), Oviedo, Spain

⁸EAP Raval Sud, Institut Català de la Salut, Red GEDAPS, Primary Care, Unitat de Suport a la Recerca (IDIAP – Fundació Jordi Gol), Barcelona, Spain

⁹Department of Endocrinology and Nutrition, Hospital Universitario Cruces – BioCruces Bizkaia – UPV-EHU, Baracaldo, Barcelona, Spain

¹⁰Diabetes Unit, Hospital Arnau de Vilanova, Valencia, Spain

¹¹Genomic Studies and Genetic Diagnosis Unit, Fundación de Investigación del Hospital Clínico de Valencia – INCLIVA, Valencia, Spain

¹²Department of Applied Mathematics, University of Málaga, Málaga, Spain

¹³Research Unit on Lipids and Atherosclerosis, Sant Joan University Hospital, Rovira i Virgili University, IISPV, Reus, Spain

¹⁴Metabolomics Platform, Universitat Rovira i Virgili, IISRV, Reus, Spain

¹⁵Biosfer Teslab, Reus, Spain

Valdés Sergio and Doulatram-Gamgaram Viyey have contributed equally to this work and share first authorship.

Josep Ribalta and Gemma Rojo-Martínez have contributed equally to this work and share last authorship.

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Correspondence

Sergio Valdés, Department of Endocrinology and Nutrition, Hospital Universitario Regional de Málaga Plaza del Hospital Civil s/n, 29009 Málaga, Spain.

Email: sergio.valdes@hotmail.es

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Abstract

Background and Aims: We aimed to assess the associations of exposure to air pollutants and standard and advanced lipoprotein measures, in a nationwide sample representative of the adult population of Spain.

Methods: We included 4647 adults (>18years), participants in the national, cross-sectional, population-based [di@bet.es](https://online.library.wiley.com/doi/10.1111/cei.14101) study, conducted in 2008–2010. Standard lipid measurements were analysed on an Architect C8000 Analyzer (Abbott Laboratories SA). Lipoprotein analysis was made by an advanced ¹H-NMR lipoprotein test (Liposcale®). Participants were assigned air pollution concentrations for particulate matter <10 µm (PM₁₀), <2.5 µm (PM_{2.5}) and nitrogen dioxide (NO₂), corresponding to the health examination year, obtained by modelling combined with measurements taken at air quality stations (CHIMERE chemistry-transport model).

Results: In multivariate linear regression models, each IQR increase in PM₁₀, PM_{2.5} and NO₂ was associated with 3.3%, 3.3% and 3% lower levels of HDL-c and 1.3%, 1.4% and 1.1% lower HDL particle (HDL-p) concentrations (*p* < .001 for all associations). In multivariate logistic regression, there was a significant association between PM₁₀, PM_{2.5} and NO₂ concentrations and the odds of presenting low HDL-c (<40 mg/dL), low HDL-p (<p25) and higher LDL particle (LDL-p) concentrations (≥p75). In subgroup analyses there were stronger associations between PM₁₀ and NO₂ and low HDL-p in men (*p* for interaction .008 and .034), and between NO₂ and low HDL-p in individuals with obesity (*p* for interaction .015).

Conclusions: Our study shows an association between the exposure to air pollutants and blood lipids in the general population of Spain, suggesting a link to atherosclerosis.

KEYWORDS

air pollution, cholesterol, HDL, lipids, lipoprotein, triglyceride

1 | BACKGROUND

The World Health Organization (WHO) has identified air pollution as the largest single environmental health risk worldwide, with outdoor air pollution accounting for more than 4.2 million deaths every year.¹ Nearly half of outdoor air pollution-related premature deaths are due to ischaemic heart disease and stroke.¹ Air pollution can impact the cardiovascular system through a number of mechanisms, including endothelial dysfunction, systemic and pulmonary oxidative stress and inflammation, autonomic nervous system dysfunction and epigenetic changes.^{2–5} Alterations in lipid metabolism could be another potential pathway in the association between air pollution and arteriosclerosis. In this regard, several previous epidemiological studies have explored the relationships between the exposure to several ambient air pollutants and blood lipid levels and the presence of

dyslipidaemia.^{6–24} However, most of this evidence has relied on standard lipid measures. Standard lipid panels measure the cholesterol or triglyceride content (in concentration per decilitre) carried by each lipoprotein class, rather than the numbers of these particles. In contrast, NMR spectroscopy provides a direct assessment of the number and size of lipoprotein particles, providing additional information about cardiovascular disease (CVD) risk.^{25–29}

Accordingly, in the present study, we aimed to assess the associations between the exposure to air pollutants [particles with an aerodynamic diameter of less than 10 microns (PM₁₀), particles with an aerodynamic diameter of less than 2.5 microns (PM_{2.5}) and nitrogen dioxide (NO₂)], with both the blood lipid levels of a standard lipid profile as well as the particle concentrations of lipoproteins, in a nationwide sample representative of the adult population of Spain.

2 | METHODS

2.1 | Study design, setting and population

The Di@bet.es study is a national, cross-sectional, population-based survey conducted in 2008–2010.³⁰ A cluster sampling design was used to select participants to form a representative random sample of the Spanish population. One hundred health centres or their equivalent from all around the country were selected at random, with a probability for selection proportional to their target population size, after which 100 individuals aged ≥ 18 years were randomly selected from each health centre. Of the eligible adults, 55.8% came for examination, of which 9.9% were excluded (institutionalized, severe disease, pregnancy or recent delivery), resulting in a final sample of 5072 individuals.

The present study focuses on 4647 individuals (92% of the study sample), in whom complete data on concentrations of air pollutants and blood lipids were available for analyses.

This research was carried out in accordance with the Declaration of Helsinki of the World Medical Association.³¹ Written informed consent was obtained from all the participants. The study was approved by the Ethics and Clinical Investigation Committee of the Hospital Regional Universitario de Málaga (Málaga, Spain) in addition to other regional ethics and clinical investigation committees all over Spain.

2.2 | Variables and procedures

The participants were invited to attend a single examination visit at their health centre. Information was collected by means of an interviewer-administered structured questionnaire, followed by a physical examination and blood sampling. Information on age, gender, educational level (none/basic/high school/college), smoking habit (current, former or never smokers) and alcohol intake ($<30/30\text{--}60/>60$ servings per month), was obtained by questionnaire. Food consumption was determined by a food frequency questionnaire and adherence to the Mediterranean diet was estimated by an adaptation of a 14-item Mediterranean diet score (MedScore).³² The level of daily physical activity was estimated by the short form of the International Physical Activity Questionnaire (IPAQ).³³ Weight and height were measured and the body mass index (BMI) was calculated as weight (kg)/height (m)². Medical history and medications were also recorded.

Blood samples, obtained in fasting conditions, were immediately centrifuged and the serum was frozen until

analysis. Samples were managed by the biochemistry laboratory of the Hospital Regional Universitario de Málaga, the IBIMA Biobank and by the CIBERDEM Biorepository (IDIBAPS Biobank).

2.3 | Lipid measurements

Standard lipid measurements Serum levels (mg/dL) of triglycerides (TG), total cholesterol (TC) and high-density lipoprotein cholesterol (HDL-c) were measured on an Architect C8000 Analyzer (Abbott Laboratories SA). Low-density lipoprotein cholesterol (LDL-c) was estimated by the Friedewald formula. 'High total cholesterol' was defined as $TC \geq 240$ mg/dL; 'high LDL-c' was defined as $LDL-c \geq 160$ mg/dL; 'low HDL-c' was defined as $HDL-c < 40$ mg/dL; and hypertriglyceridaemia was defined as $TG \geq 200$ mg/dL.³⁴

Lipoprotein analysis Lipoprotein analysis was performed by using Liposcale® Test (CE, ISO 13.485 approved), an advanced lipoprotein test based on two-dimensional (2D) diffusion-ordered ¹H-NMR (Proton nuclear magnetic resonance) spectroscopy.³⁵ Before ¹H-NMR analysis, 200 μ L of serum were diluted with 50 μ L deuterated water and 300 μ L of 50 mM phosphate buffer solution at pH 7.4. ¹H-NMR spectra were recorded at 306K on a Bruker Avance III 600 spectrometer operating at a proton frequency of 600.20 MHz. The methyl signal was deconvoluted by using Lorentzian functions to determine the lipid concentration of the main lipoprotein classes (VLDL, LDL and HDL), and their size associated diffusion coefficients. Then, the lipid concentrations were combined with their associated particle volume in order to quantify the number of particles required to transport the measured lipid concentration of each lipoprotein subclass calculating the particle concentrations of VLDL (VLDL-p), LDL (LDL-p) and HDL (HDL-p). The variation coefficients for particle number were between 2% and 4%, and for the particle sizes were lower than 0.3%. These analyses were performed at the Biosfer Teslab facilities. We classified participants as having normal or abnormal levels of LDL-p, VLDL-p and HDL-p, using the equivalents of the 75th (LDL-p and VLDL-p) and 25th (HDL-p) percentiles (p) of particle concentrations within our study population.

2.4 | Exposure assessment

Modelled mean annual PM₁₀, PM_{2.5} and NO₂ concentrations in Spain for the period 2008–2010 were calculated with the CHIMERE chemistry-transport model.³⁶ This model calculates the concentration of gaseous species and both inorganic and organic aerosols of primary and

secondary origin, including primary particulate matter, mineral dust, sulphate, nitrate, ammonium, secondary organic species and water. This model has been broadly evaluated in Spain by comparison with measured air pollutants at a large set of monitoring sites.^{37,38} The model was applied to a domain covering the Iberian Peninsula at a horizontal resolution of $0.1 \times 0.1^\circ$. The modelled concentrations were corrected with observed values, by considering a methodology described by Martín et al.³⁹ in which (1) a bias is calculated with regard to the observations in the Spanish air quality network of monitoring sites, (2) these biases are spatially interpolated using a kriging methodology to obtain a gridded bias and (3) this gridded bias is applied to the modelled concentration grid. This methodology considers a different bias grid for rural and urban sites that are then combined and weighted by population density. We assigned the average annual exposure to air pollutants corresponding to the health examination year of each participant by interpolating the estimated concentrations to the centroid of their residential postal codes.

Data on mean annual temperature ($^\circ\text{C}$) from each municipality of residence were obtained from the Spanish National Meteorological Agency.⁴⁰

2.5 | Statistical analysis

We applied linear regression models to assess associations between air pollutant and lipid measurements, which were log transformed to normalize distributions and also to limit the influence of extreme values. Association estimates were presented as percent changes with corresponding 95% confidence intervals (calculated by $100 \times [\exp(b) - 1]$), per each interquartile range (IQR) increase in air pollutant concentrations which equated to $7.7 \mu\text{g}/\text{m}^3 \text{PM}_{10}$, $4.8 \mu\text{g}/\text{m}^3 \text{PM}_{2.5}$ and $12.1 \mu\text{g}/\text{m}^3 \text{NO}_2$.

We also used logistic regression models to investigate the associations of ambient air pollutants with high total cholesterol ($\text{TC} \geq 240 \text{ mg}/\text{dL}$); high LDL-c ($\text{LDL-c} \geq 160 \text{ mg}/\text{dL}$); low HDL-c ($\text{HDL-c} < 40 \text{ mg}/\text{dL}$); and hypertriglyceridaemia ($\text{TG} \geq 200 \text{ mg}/\text{dL}$) as defined by the standard lipid measurements, and with high LDL-p and VLDL-p concentrations ($\geq p75$), and low HDL-p concentrations ($< p25$), measured by the Liposcale[®] test. These results are presented as odds ratios (ORs) with corresponding 95% CIs again per each IQR increase in air pollutants.

All these models were controlled for possible confounders such as age, sex, BMI, education level, smoking status, alcohol intake, municipality population, MedScore, IPAQ, mean ambient temperature of the municipality and lipid lowering medication.

In addition, we investigated potential effect modification of the associations by sex (male/female), age

(<40/40–60 or ≥ 60 years) and BMI (<30 or $\geq 30 \text{ kg}/\text{m}^2$). Each potential modifier was examined in a separate model by adding an interaction term.

All the statistical analyses were performed with IBM SPSS statistics 23.0. Reported *p* values were based on two-sided tests with statistical significance set at .05. Bonferroni correction for multiple comparisons was applied at a level of $\alpha = .05/7 = .007$.

3 | RESULTS

A total of 4647 individuals were included in the analysis (Table 1). The sample was composed of 1976 men (42.5%) and 2671 women (57.5%). Mean age of the population was 50.5 ± 17.0 years (range: 18–93 years). The characteristics of the study population followed a distribution as expected in the Spanish general population. Distributions of standard lipids and NMR particle concentrations are also displayed in the table.

Table 2 summarizes residential estimates of outdoor air pollution concentrations assigned to the study participants in the year of examination. The median (25th–75th percentile) PM_{10} , $\text{PM}_{2.5}$ and NO_2 exposure levels were 23.7 (19.6–27.3), 12.2 (10.5–15.3) and 16.6 (12.5–24.6) $\mu\text{g}/\text{m}^3$, respectively. Most values were within the current European Ambient Air Quality Directive target values (Directive 2008/50/EC).⁴¹

Table 3 shows the results of the linear correlations between air pollutant concentrations and standard lipid biomarkers (TC, LDL-c, HDL-c and TG) and NMR particle concentrations (LDL-p, VLDL-p and HDL-p), in crude, and multivariate adjusted linear regression models. Greater air pollutant exposures were associated with lower TC and LDL-c concentrations in crude models, but these associations were attenuated after multivariate adjustment. This trend was not observed by NMR analyses. In fact the association between air pollutants and LDL-p although not significant tended to be positive. The three pollutants tested showed a strong and highly significant negative association with HDL-c concentrations in both crude and multivariate models. In the fully adjusted model, an IQR increase in PM_{10} , $\text{PM}_{2.5}$ and NO_2 was associated with 3.3%, 3.3% and 3% lower levels of HDL-c, respectively. The same trend was observed regarding the association between contaminants and the concentration of HDL particles measured by NMR. In the multivariate analysis, each IQR increase in PM_{10} , $\text{PM}_{2.5}$ and NO_2 was associated with 1.3%, 1.4% and 1.1% lower HDL-p. $\text{PM}_{2.5}$ concentrations also showed a positive association with TG levels in multivariate analyses.

Table 4 shows the crude and multivariate-adjusted ORs for presenting lipid abnormalities in the standard lipid

TABLE 1 Clinical characteristics and lipid values of the study sample ($n=4647$).

| | % | Mean \pm SD | Range |
|---|------|-----------------|-----------|
| Age (years) | | 50.5 \pm 17.0 | 18–93 |
| Women | 57.5 | | |
| Smoking | | | |
| Current | 23.8 | | |
| Former | 26.0 | | |
| Never | 50.2 | | |
| Alcohol intake (servings-month) | | | |
| <30 | 73.9 | | |
| 30–60 | 14.9 | | |
| >60 | 11.3 | | |
| Municipality population | | | |
| <10,000 | 18.4 | | |
| 10,000–50,000 | 27.8 | | |
| >50,000 | 53.8 | | |
| Education level | | | |
| No studies | 12.8 | | |
| Basic | 47.6 | | |
| High school-college | 39.6 | | |
| BMI (kg/m ²) | | 28.0 \pm 5.2 | 12.2–61.3 |
| Med diet score | | 7.8 \pm 1.8 | 1–13 |
| Physical activity (IPAQ) | | | |
| Low | 42.0 | | |
| Medium | 34.5 | | |
| High | 23.5 | | |
| Lipid lowering medication | 13.3 | | |
| Standard lipid biomarkers (mg/dL) | | | |
| TC | | 196 \pm 40 | 58–395 |
| LDL-C | | 105 \pm 30 | 21–254 |
| HDL-C | | 52 \pm 13 | 7–14 |
| TG | | 121 \pm 88 | 17–2095 |
| Particle concentration (¹ H-NMR) (nmol/L) | | | |
| LDL-p | | 1392 \pm 278 | 358–2702 |
| VLDL-p | | 53 \pm 40 | 12–619 |
| HDL-p | | 28 \pm 5 | 6–56 |

Abbreviations: HDL-C, high-density lipoprotein cholesterol; HDL-p, HDL particles; LDL-C, low-density lipoprotein cholesterol; LDL-p, LDL particles; TC, total cholesterol; TG: triglycerides; VLDL-p, VLDL particles; ¹H-NMR, proton nuclear magnetic resonance.

TABLE 2 Descriptive statistics for air pollutants concentrations ($\mu\text{g}/\text{m}^3$) in the study sample.

| Pollutant | Percentile | | | | | IQR | Mean | Minimum | Maximum |
|-------------------|------------|------|------|------|------|------|------|---------|---------|
| | 5th | 25th | 50th | 75th | 95th | | | | |
| PM ₁₀ | 14.4 | 19.6 | 23.7 | 27.3 | 33.1 | 7.7 | 23.2 | 4.0 | 42.3 |
| PM _{2.5} | 8.1 | 10.5 | 12.2 | 15.3 | 20.0 | 4.8 | 12.7 | 3.4 | 22.3 |
| NO ₂ | 6.6 | 12.5 | 16.6 | 24.6 | 50.3 | 12.1 | 20.2 | 3.6 | 51.4 |

Abbreviations: IQR, interquartile range; NO₂, nitrogen dioxide; PM₁₀, particles with an aerodynamic diameter of less than 10 microns; PM_{2.5}, particles with an aerodynamic diameter of less than 2.5 microns.

profile and NMR, per each IQR increase in air pollutant concentrations. Replicating what was seen in the linear model, there was a strong negative association between the three air pollutants tested, and HDL-c. In the fully adjusted model, the odds of presenting low HDL-c (<40 mg/dL), were 1.36 (95% CI 1.21–1.54) $p < .001$, 1.41 (95% CI 1.25–1.59) $p < .001$ and 1.31 (95% CI 1.18–1.44) $p < .001$ per each IQR increase in PM₁₀, PM_{2.5} and NO₂. Again, this same trend was observed for the association between air pollutants and HDL-p, with multivariate ORs for presenting low HDL-p (<p25) of 1.23 (95% CI 1.09–1.37) $p < .001$, 1.23 (95% CI 1.10–1.37) $p < .001$ and 1.18 (95% CI 1.07–1.29) $p = .001$ per each IQR increase in PM₁₀, PM_{2.5} and NO₂, respectively. In the logistic regression models, there was also a positive association between PM₁₀ and PM_{2.5} concentrations and the odds of presenting higher LDL-p, with multivariate ORs for presenting LDL-p concentrations $\geq p75$ of 1.24 (95% CI 1.11–1.39) $p < .001$ and 1.16 (95% CI 1.04–1.29) $p = .007$, per each IQR increase, respectively.

In the subgroup analysis, we found a significant interaction between sex and the association between PM₁₀ and NO₂ and low HDL-p (p for interaction .008 and .034, respectively), with stronger associations in men. We also found a stronger association between NO₂ and low HDL-p in individuals with obesity (p for interaction .015) (Figure 1).

4 | DISCUSSION

In this nationwide sample representative of the adult population of Spain we found significant associations between exposure to various air pollutants and several standard and novel blood lipoprotein measures, pointing to a pro-atherogenic lipid profile in subjects exposed to a higher degree of pollution. In particular we found strong negative associations between PM₁₀, PM_{2.5} and NO₂ exposures, and both HDL-c and HDL-p concentrations both in linear and in logistic multivariate regression models. Additionally, we found a significant association between exposure to PM₁₀ and PM_{2.5}, and

TABLE 3 Associations between air pollutants (per IQR concentrations increase) and blood lipid levels.

| Standard lipid biomarkers | | | | | | | | | | | | |
|--------------------------------|--------------------------|----------|--|--------------------------|----------|--|--------------------------|----------|--|-----------------------|----------|--|
| | TC | | | LDL-C | | | HDL-C | | | TG | | |
| | % change (95% CI) | <i>p</i> | | % change (95% CI) | <i>p</i> | | % change (95% CI) | <i>p</i> | | % change (95% CI) | <i>p</i> | |
| PM ₁₀ | | | | | | | | | | | | |
| Crude | -1.7 (-2.5, -1.0) | <.001 | | -1.6 (-2.7, -0.5) | .004 | | -3.2 (-4.1, -2.3) | <.001 | | 0.4 (-1.4, 2.0) | .667 | |
| Multivariate | -1.1 (-1.9, -0.2) | .013 | | -0.3 (-1.6, 0.9) | .590 | | -3.3 (-4.3, -2.3) | <.001 | | 2.0 (0.1, 3.9) | .037 | |
| PM _{2.5} | | | | | | | | | | | | |
| Crude | -1.1 (-1.9, -0.3) | .01 | | -0.8 (-2.0, 0.4) | .190 | | -3.0 (-4.0, -2.1) | <.001 | | 1.0 (-1.0, 2.9) | .323 | |
| Multivariate | -0.6 (-1.4, 0.3) | .215 | | -0.3 (-1.0, 1.5) | .628 | | -3.3 (-4.3, -2.3) | <.001 | | 2.8 (0.9, 4.7) | .003 | |
| NO ₂ | | | | | | | | | | | | |
| Crude | -1.8 (-2.5, -1.2) | <.001 | | -1.8 (-2.7, -0.9) | <.001 | | -2.7 (-3.4, -1.9) | <.001 | | -1.5 (-2.9, 0.0) | .053 | |
| Multivariate | -1.4 (-2.2, -0.7) | <.001 | | -0.9 (-1.9, 0.2) | .118 | | -3.0 (-3.8, -2.2) | <.001 | | -0.5 (-1.2, 2.1) | .570 | |
| Particle concentration (H-NMR) | | | | | | | | | | | | |
| | LDL-p | | | VLDL-p | | | HDL-p | | | | | |
| | % change (95% CI) | <i>p</i> | | % change (95% CI) | <i>p</i> | | % change (95% CI) | <i>p</i> | | % change (95% CI) | <i>p</i> | |
| PM ₁₀ | | | | | | | | | | | | |
| Crude | 0.2 (-0.6, 0.9) | .703 | | 1.1 (-1.0, 3.2) | .324 | | -0.7 (-1.3, 0.0) | .049 | | | | |
| Multivariate | 0.6 (-0.3, 1.4) | .205 | | 1.9 (-0.3, 4.1) | .091 | | -1.3 (-2.0, -0.6) | <.001 | | | | |
| PM _{2.5} | | | | | | | | | | | | |
| Crude | 0.4 (-0.4, 1.2) | .364 | | 1.0 (-1.3, 3.2) | .391 | | -0.8 (-1.5, -0.1) | .018 | | | | |
| Multivariate | 0.6 (-0.2, 1.5) | .159 | | 1.9 (-0.3, 4.1) | .089 | | -1.4 (-2.1, -0.7) | <.001 | | | | |
| NO ₂ | | | | | | | | | | | | |
| Crude | 0.6 (0.0, 1.2) | .068 | | 0.3 (-1.3, 2.0) | .703 | | -0.5 (-1.1, 0.0) | .040 | | | | |
| Multivariate | 0.5 (-0.2, 1.3) | .155 | | 1.8 (-0.1, 3.6) | .059 | | -1.1 (-1.6, -0.5) | <.001 | | | | |

Note: % changes and *p* values calculated by linear regression per interquartile range (IQR) increase in air pollutants concentrations (PM₁₀: 7.7 µg/m³, PM_{2.5}: 4.8 µg/m³, NO₂: 12.1 µg/m³). Multivariate model: adjusted to age, sex, BMI, smoking status, alcohol intake, education level, MedScore, IPAQ, municipality population, ambient temperature and lipid lowering medication. In bold: % changes with *p* values < .007 (alpha .05 corrected by Bonferroni = .05/7).

Abbreviations: HDL-C, high-density lipoprotein cholesterol; HDL-p, HDL particles; LDL-C, low-density lipoprotein cholesterol; LDL-p, LDL particles; NO₂, nitrogen dioxide; PM₁₀, particles with an aerodynamic diameter of less than 10 microns; PM_{2.5}, particles with an aerodynamic diameter of less than 2.5 microns; TC, total cholesterol; VLDL-p, VLDL particles; H-NMR, proton nuclear magnetic resonance.

TABLE 4 Odd ratios (OR) for presenting lipid abnormalities per interquartile range (IQR) increase in air pollutants concentrations.

| | Standard lipid biomarkers | | | | | | | | | | | |
|-------------------------|--|------------------|----------|------------------------|-----------|----------|--------------------|------------------|----------|--------------------------------|-----------|----------|
| | TC \geq 240 mg/dL | | | LDL-C \geq 160 mg/dL | | | HDL-C $<$ 40 mg/dL | | | Triglycerides \geq 200 mg/dL | | |
| | OR | 95% CI | <i>p</i> | OR | 95% CI | <i>p</i> | OR | 95% CI | <i>p</i> | OR | 95% CI | <i>p</i> |
| PM₁₀ | | | | | | | | | | | | |
| Crude | 0.94 | 0.85–1.05 | .282 | 0.99 | 0.82–1.20 | .952 | 1.29 | 1.17–1.43 | .000 | 0.96 | 0.85–1.08 | .489 |
| Multivariate | 1.02 | 0.90–1.16 | .718 | 1.10 | 0.87–1.37 | .428 | 1.36 | 1.21–1.54 | .000 | 0.97 | 0.83–1.12 | .655 |
| PM_{2.5} | | | | | | | | | | | | |
| Crude | 1.00 | 0.89–1.13 | .958 | 1.07 | 0.87–1.31 | .521 | 1.33 | 1.20–1.47 | .000 | 0.94 | 0.82–1.07 | .353 |
| Multivariate | 1.07 | 0.94–1.22 | .314 | 1.16 | 0.92–1.46 | .198 | 1.41 | 1.25–1.59 | .000 | 0.95 | 0.82–1.11 | .542 |
| NO₂ | | | | | | | | | | | | |
| Crude | 0.93 | 0.85–1.02 | .139 | 0.97 | 0.83–1.14 | .975 | 1.23 | 1.14–1.32 | .000 | 0.91 | 0.82–1.01 | .073 |
| Multivariate | 0.99 | 0.88–1.11 | .853 | 1.11 | 0.91–1.36 | .285 | 1.31 | 1.18–1.44 | .000 | 0.96 | 0.84–1.10 | .558 |
| | Particle concentration (¹ H-NMR) | | | | | | | | | | | |
| | LDL-p \geq p75 | | | VLDL-p \geq p75 | | | HDL-p $<$ p25 | | | | | |
| | OR | 95% CI | <i>p</i> | OR | 95% CI | <i>p</i> | OR | 95% CI | <i>p</i> | OR | 95% CI | <i>p</i> |
| PM₁₀ | | | | | | | | | | | | |
| Crude | 1.15 | 1.06–1.26 | .002 | 1.03 | 0.94–1.12 | .578 | 1.12 | 1.03–1.23 | .011 | | | |
| Multivariate | 1.24 | 1.11–1.39 | .000 | 1.08 | 0.96–1.21 | .188 | 1.23 | 1.09–1.37 | .000 | | | |
| PM_{2.5} | | | | | | | | | | | | |
| Crude | 1.12 | 1.02–1.23 | .016 | 1.02 | 0.93–1.12 | .631 | 1.13 | 1.03–1.24 | .007 | | | |
| Multivariate | 1.16 | 1.04–1.29 | .007 | 1.09 | 0.97–1.21 | .144 | 1.23 | 1.10–1.37 | .000 | | | |
| NO₂ | | | | | | | | | | | | |
| Crude | 1.08 | 1.01–1.16 | .023 | 0.99 | 0.92–1.06 | .741 | 1.10 | 1.03–1.17 | .007 | | | |
| Multivariate | 1.11 | 1.02–1.22 | .018 | 1.05 | 0.95–1.15 | .334 | 1.18 | 1.07–1.29 | .001 | | | |

Note: ORs, 95% CI and *p* values were calculated by logistic regression per interquartile range (IQR) increase in air pollutants concentrations (PM₁₀: 7.7 $\mu\text{g}/\text{m}^3$, PM_{2.5}: 4.8 $\mu\text{g}/\text{m}^3$, NO₂: 12.1 $\mu\text{g}/\text{m}^3$). In bold: % changes with *p* values $<$.007 (alpha .05 corrected by Bonferroni = .05/7). Multivariate model: adjusted to age, sex, BMI, smoking status, alcohol intake, education level, MedScore, IPAQ, municipality population, ambient temperature and lipid lowering medication.

Abbreviations: HDL-C, high-density lipoprotein cholesterol; HDL-p, HDL particles; LDL-C, low-density lipoprotein cholesterol; LDL-p, LDL particles; NO₂, nitrogen dioxide; PM₁₀, particles with an aerodynamic diameter of less than 10 microns; PM_{2.5}, particles with an aerodynamic diameter of less than 2.5 microns; TC, total cholesterol; TG, triglycerides; VLDL-p, VLDL particles; ¹H-NMR, proton nuclear magnetic resonance.

higher LDL-p concentrations (\geq p75), although the associations between air pollutants and the standard LDL-c measures were null.

These findings are in line with previous studies suggesting that air pollution might negatively impact blood lipids in the general population,^{6–24} and may be a potential contributor to air pollution-related CVD risk. In fact, low HDL-c is a well-established atherosclerotic CVD risk factor,³⁴ whereas lower HDL and higher LDL particle numbers have also been associated with increasing cardiovascular events.^{42–44}

Interestingly, although the relationship of the exposure to different air pollutants with standard lipid biomarkers has been widely studied, to the best of our knowledge, only two previous studies have assessed associations with particle concentration measures. Bell

et al.⁹ examined the relationship between air pollution and both HDL-c and HDL-p, in 6654 men and women free of prevalent clinical CVD, participants in the Multi-Ethnic Study of Atherosclerosis Air Pollution study (MESA Air). A 5 $\mu\text{g}/\text{m}^3$ higher PM_{2.5} was associated with lower HDL-p ($-0.64 \mu\text{mol}/\text{L}$ [95% CI $-1.01, -0.26$]), but not HDL-c ($-0.05 \text{ mg}/\text{dL}$ [95% CI $-0.82, 0.71$]). McGuinn et al.¹³ studied linear associations to estimate change in lipoprotein levels with each $\mu\text{g}/\text{m}^3$ increase in annual average PM_{2.5} in 6587 patients who had a cardiac catheterization in Duke University between 2001 and 2010 (CATHGEN study). The percent change from the mean outcome level was 2.00% (95% CI 1.38%, 2.64%) for total LDL-p. However, the associations between air pollution and HDL particle concentrations in this study were inconsistent. This is in contrast to the findings

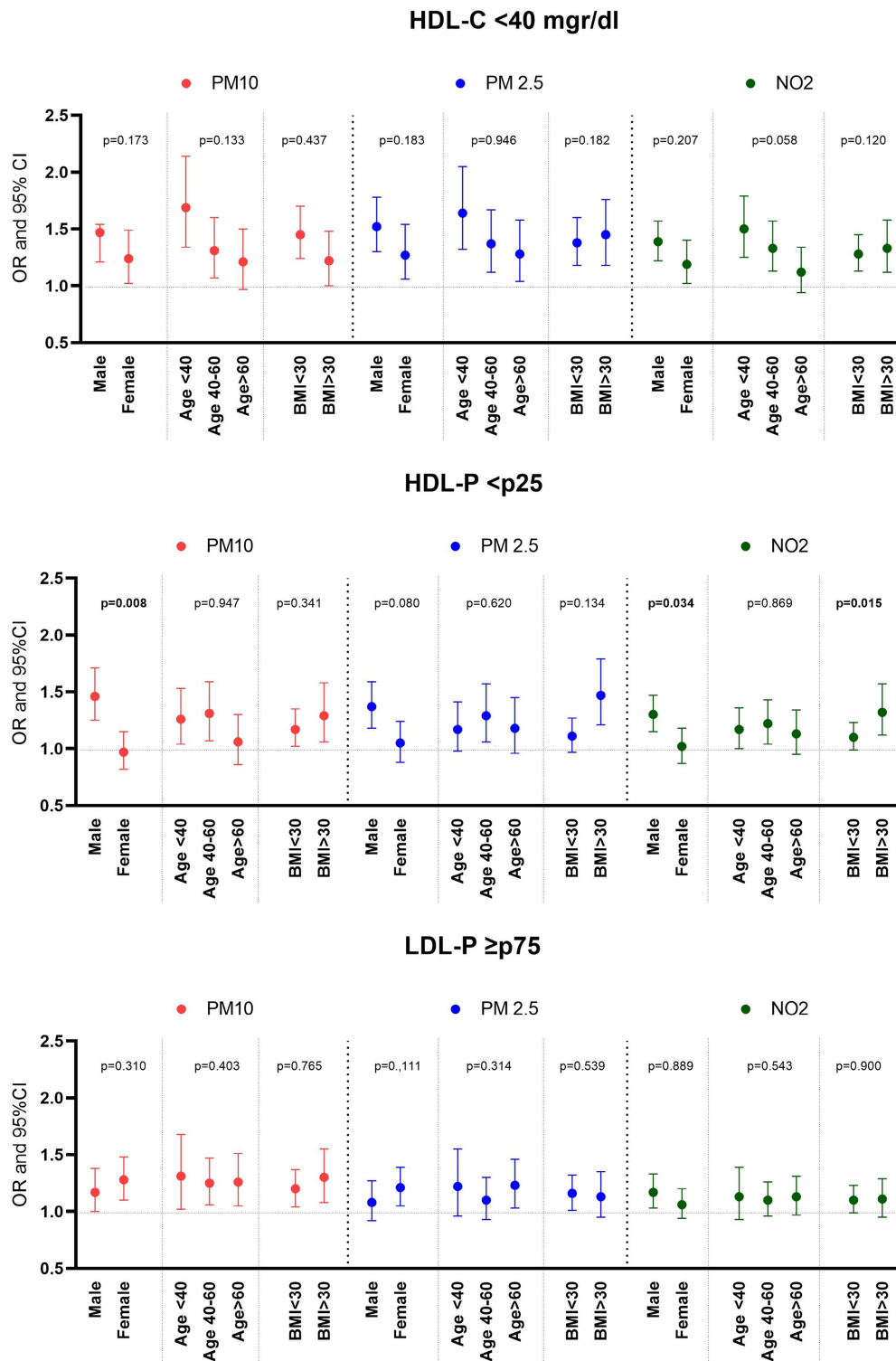


FIGURE 1 Logistic regression analyses between $PM_{2.5}$, PM_{10} and NO_2 exposures and HDL-c <40 mg/dL, HDL-p <p25 and LDL-p \geq p75 in different population subgroups.

reported by Bell et al. in the MESA Air study⁹ and to our results in [di@bet.es](https://doi.org/10.1111/cei.14101), which indicate strong negative associations between air pollutants and HDL markers. Of note, the CATHGEN study sample consisted of patients who underwent cardiac catheterization, who may

represent a highly selective population, while the MESA Air and the [di@bet.es](https://doi.org/10.1111/cei.14101) study populations may be more generalizable to the general background population.

Although the observed differences in blood lipids found in our study may seem small, it is of note that, as

pointed out by Bell et al.,⁹ the magnitude of the decrease of HDL-c and HDL-p we have observed (between -3.0% and -3.3% decrease in HDL-c and -1.1% to -1.4% decrease in HDL-p per IQR increase in air pollutants concentration), can be compared to the effect of smoking on these measurements as observed in smoking cessation studies (2.4 mg/dL and 1.0 $\mu\text{mol/L}$ change on HDL-c and HDL-p, respectively).⁴⁵

On the other hand, exposure to air pollution has also been shown to induce the development of dysfunctional HDL resulting in alteration of its atheroprotective capacities,^{46,47} which may contribute to atherosclerosis progression beyond the total lipoprotein concentrations measured in our study.

The apparently counterintuitive negative association between air pollutants and TC observed is of small magnitude and could be explained by the contribution of HDL-c to the TC levels, which was significantly reduced when associated with higher exposures. In line with this, no associations with LDL-c levels in the multivariate analyses were found. This apparently benign effect of air pollutants in the standard LDL-c measures was however accompanied by a significant association between higher exposures to PM₁₀ and PM_{2.5}, and higher LDL-p concentrations (≥ 75). LDL-p has been found to be a better predictor of CVD than LDL-c, especially in individuals with LDL-p/LDL-c discordance.^{28,29,48} So again, these associations could be regarded as pro-atherogenic. In subgroup analyses, we found stronger associations between PM₁₀ and NO₂ and low HDL-p in men, and between NO₂ and low HDL-p in individuals with obesity, suggesting that these subgroups may be more susceptible. A stronger susceptibility in overweight-obese individuals is consistent with Sørensen et al.,⁶ Yang et al.,¹⁰ Kim et al.,¹² Mao et al.,¹⁶ Zhang et al.²² and Kim et al.¹⁹ among others. Both air pollution exposure and overweight/obesity are associated with higher systemic inflammation^{2,49} which may explain an interplay of these factors. The modification effects of sex in previous studies have been mixed and harder to interpret.

The underlying mechanisms explaining the effect of air pollutants on lipids have not been fully elucidated. The main hypothesis is that the oxidative stress and systemic inflammation caused by inhaled air pollution could induce adverse lipid metabolism and lipid oxidation.^{50,51} Air pollutants have also been found to cause DNA methylation of genes related to lipid metabolism.⁵² Further studies are warranted to clarify the full spectrum of these mechanisms.

Our study has several strengths, including the large population-based design and the inclusion of advanced lipoprotein measures, as well as other extensive individual-level data of clinical, demographic and lifestyle variables, which allowed us to perform a complete multivariate

adjustment of the data. Our nationwide perspective, allows us to extrapolate our results more widely than local or regional studies, increasing the public health implications of the findings.

The limitations of our study include its observational cross-sectional nature; so that we cannot establish causal associations or exclude residual confounding in the relation between air pollutants and lipids. Also, we used ambient outdoor measurements modelled at the residential addresses of the participants as a proxy for exposure to air pollution, whereas no other relevant information such as time-activity patterns, proximity to main roads, occupational exposures or personal monitoring data was available. Exposure measurement error is possible when using modelled pollutant levels and this could, in fact, have attenuated our effect estimates. Finally, our exposure models were developed based on yearly exposures to air pollutants, whereas more refined measures to look at different lags were not available.

5 | CONCLUSIONS

In summary, our study, in keeping with previous data, suggests a deleterious effect of the exposure to air pollutants on blood lipids, in the general population of Spain. Our results reinforce the need for improving air quality as much as possible to decrease the risk of atherosclerosis in our population, as the lipid changes observed in our study may be a potential contributor to air pollution-related CVD risk.

AUTHOR CONTRIBUTIONS

Conception and design: G.R.M and J.R. Acquisition of epidemiological data: C.M.A., E.G.E., S.G.S., A.C.P., L.C., E.D., E.M., J.F.N., S.G., J.G., F.J.C., S.V. and G.R.M. tandard biochemical samples management: W.O.B. and G.R.M. Liposcale® samples management and analyses: J.R., N.A. and M.G. Air pollution modelling: M.G.V., J.L.G., M.T., V.G. and F.M.L. Creation of new software used in the work: J.L.G.G and G.A.V. Analysis and interpretation of data: S.V. J.C.V. and G.R.M. Drafting the article: S.V., V.K.D.G. and G.R.M. All authors revised and approved the final manuscript.

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CONFLICT OF INTEREST STATEMENT

N.A. is stock owner of Biosfer Teslab and has a patent on the method for lipoprotein profiling described in the present manuscript. The other authors declare that they have no competing financial interests related to this manuscript.

ORCID

Sergio Valdés  <https://orcid.org/0000-0001-8623-4331>

Viyey Doulatram-Gamgaram  <https://orcid.org/0000-0002-0178-0040>

Cristina Maldonado-Araque  <https://orcid.org/0000-0002-3452-3489>

Eva García-Escobar  <https://orcid.org/0000-0002-0134-3473>

Joan Girbés  <https://orcid.org/0000-0002-1665-4549>

Gemma Rojo-Martínez  <https://orcid.org/0000-0003-2179-2134>

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